### The Development of CO<sub>2</sub> Plume in CO<sub>2</sub> Sequestration in the Aquifer

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#### Abstract

Geological carbon sequestration through injecting large-scale carbon dioxide (CO<sub>2</sub>) into the deep saline aquifers represents a long-term storage of  $CO_2$ . In the  $CO_2$  sequestration process, the injected  $CO_2$ is displacing water from the injection point and is expected to remain in the reservoir. Due to the nature of one phase displacing another phase in porous media, it is noted that different water saturation exists in the  $CO_2$  plume during the displacement. Water distribution in the plume will affect the size of the plume subsurface. Furthermore, the gravitational segregation between CO<sub>2</sub> and water will cause overridingtonguing during the injection and impact the shape of plume. To better understand the CO<sub>2</sub> movement underground and development of  $CO_2$  plume, it is necessary to take the two-phase flow and gravity force effects into account when evaluating CO<sub>2</sub> displacing water. The displacement of water by injecting CO<sub>2</sub> is not a piston-like process in aquifer. Because water is the wetting phase and CO<sub>2</sub> is the non-wetting phase when two phases flow in reservoir, water occupies the surface of matrix and small pores while CO<sub>2</sub> resides in large pores and centers of pores. As a result, various water saturations distribute behind CO<sub>2</sub> front during the displacement. The distribution is a function of fluid and rock properties, fluid-rock interaction, and injection operation. In this study, these factors are considered when developing new models to predict CO<sub>2</sub> plume evolution during injection. Mass conservation, multiphase flow, and equation-of-states are applied in the derivation of the models, which guarantees a rigorous approach in the investigation. The modeling results indicate that CO<sub>2</sub> does not displace water completely away from the plume. The shape of the CO<sub>2</sub> front is controlled by the relative permeability of two phases and capillary pressure. Water saturation profile from CO<sub>2</sub> injecting point to the displacement front shows that water saturation behind the CO<sub>2</sub> front increases outwardly, and the change in saturation is non-linear. The injection rate impacts the sharpness of the CO<sub>2</sub> front, thus leads to different gas plume sizes for same injection volume. The outward movement of the CO<sub>2</sub> front decelerates as injection time goes on. The research illustrates that injection experiences two stages: transient and steady-state, in which the displacement behavior and the development of gas plume vary. Although the duration of transient stage is dictated by size of aquifer and is relatively short comparing with steady-state stage, its influence on the development of CO<sub>2</sub> plume cannot be neglected when selecting gas compressor horsepower and determining injection rate.

### Introduction

Greenhouse effect is one of the processes that cause the arguable global warming in recent years. When the energy from the Sun reaches the Earth, part of it is reflected back to the space while the rest is absorbed and reradiated by the Earth's atmosphere, which consists of greenhouse gases. CO<sub>2</sub> is one of the greenhouse gases that contribute to the trapping of heat that radiates from Earth. Greenhouse gases in the Earth's atmosphere are mainly CO<sub>2</sub>, methane, nitrous oxide, hydrofluorocarbons, sulfur hexafluoride, fluorinated gases, ozone (O<sub>3</sub>), and water vapor, etc. Table 1 (Haynes, 2016) shows the concentration of different gas in atmosphere. It is noted that the contribution of each greenhouse gas to the absorption of

heat depends on its concentration and its global warming potential (GWP). GWP is used to reflect how long a greenhouse gas remains in the atmosphere, on average, and how strongly it absorbs energy. Gases with a higher GWP absorb more energy, per pound, than gases with a lower GWP, and thus contribute more to warming Earth. GWP of greenhouse gases is listed in Table 2 (U.S. EPA, 2019). Although the global warming potential of  $CO_2$  is not the highest, its effect on global warming is significant due to the relatively high concentration and the increasing emission volume during last several decades. Many researches indicate that anthropogenic greenhouse gases emitting to the atmosphere are the main cause of global warming. Fig. 1 illustrates that the emission of  $CO_2$  has increased dramatically since industrial revolution (Ritchie and Roser, 2019). Fig. 2 is the corresponding change of  $CO_2$  concentration in atmosphere as a result of increase in  $CO_2$  emission.  $CO_2$  emission is from burning fossil fuels, solid waste, trees and other biological materials. Chemical reactions also emit  $CO_2$ .

Gas		Volume <sup>(A)</sup>						
Name	Formula	in ppmv <sup>(B)</sup>	mole %					
Nitrogen	$N_2$	780,840	78.084					
Oxygen	$O_2$	209,460	20.946					
Argon	Ar	9,340	0.9340					
Carbon dioxide	$CO_2$	413.32	0.041332					
Neon	Ne	18.18	0.001818					
Helium	He	5.24	0.000524					
Methane	CH <sub>4</sub>	1.87	0.000187					
Krypton	Kr	1.14	0.000114					
Not included in a	above dry atmosp	ohere:						
Water vapor <sup>(C)</sup>	H <sub>2</sub> O	0-30,000 <sup>(D)</sup>	0-3% <sup>(D)</sup>					
notes: (A) Volume fraction is equal to mole fraction for ideal gas only, (B) ppmv: parts per million by volume (C) Water vapor is about 0.25% by mass over full atmosphere (D) Water vapor strongly varies locally (Wallace et al. 2006)								

Table 1. Major constituents of dry air, by volume (Haynes, 2016)

Table 2. Global Warming Potentials and Atmospheric Lifetimes (U.S. EPA, 2019)

Gas	Atmospheric Lifetime (years)	GWP <sup>a</sup>
Carbon dioxide (CO <sub>2</sub> )	50-200	1
Methane (CH <sub>4</sub> ) <sup>b</sup>	12±3	21
Nitrous oxide (N <sub>2</sub> O)	120	310
HFC-23	264	11,700
HFC-32	5.6	650

HFC-125	32.6	2,800
HFC-134a	14.6	1,300
HFC-143a	48.3	3,800
HFC-152a	1.5	140
HFC-227ea	36.5	2,900
HFC-236fa	209	6,300
HFC-4310mee	17.1	1,300
CF <sub>4</sub>	50,000	6,500
$C_2F_6$	10,000	9,200
$C_4F_{10}$	2,600	7,000
C <sub>6</sub> F <sub>14</sub>	3,200	7,400
SF <sub>6</sub>	3,200	23,900

<sup>a</sup> 100 year time horizon

<sup>b</sup> The methane GWP includes the direct effects and those indirect effects due to the production of tropospheric ozone and stratospheric water vapor. The indirect effect due to the production of  $CO_2$  is not included.



Figure 1. Global emission of CO<sub>2</sub> has increased significantly since industrial evolution (Ritchie and Roser, 2019)



Figure 2. The change of CO<sub>2</sub> concentration in atmosphere (Ritchie and Roser, 2019)

 $CO_2$  can be removed from the atmosphere by the biological carbon cycle, or the photosynthesis, a process by which green plants and some organisms use sunlight to synthesize foods from  $CO_2$  and water. Unfortunately, the amount of  $CO_2$  can be consumed by photosynthesis is lagged behind the  $CO_2$  emission as global population increases. To reduce the concentration of  $CO_2$  in atmosphere and alleviate the global warming, two approaches are employed. One is to reduce the emission of  $CO_2$  through improving fuel efficiency, moving to clean energy or renewable energy such as wind, solar, replacing coal/oil with less carbon-intense fuel (natural gas). Another is the sequestration of  $CO_2$ . It is a long-term storage of  $CO_2$  to slow the atmosphere and marine accumulation of  $CO_2$  to mitigate global warming. In the sequestration,  $CO_2$  is captured through biological, chemical, and physical process and then transported and stored in subsurface aquifers, coal beds, depleted oil and gas reservoirs, and ocean water. The storage of  $CO_2$ underground in aquifers, coal beds, depleted oil and gas reservoirs is often referred to geological sequestration.

When selecting a site for geological sequestration, several parameters are considered: reservoir porosity, permeability, existing of fault, sealability of caprock, geometry of reservoir, formation depth, and formation thickness. Porosity and thickness affect the formation storage capacity. Permeability dictates the injection rate and migration of  $CO_2$  in formation. Faults cause the non-uniform flow or tonging phenomenon. Faults in the reservoir are the preferential path for  $CO_2$  to flow. Caprock keeps  $CO_2$  remaining in formation for a long geological time after the injection. Once injected, the  $CO_2$  plume will rise via buoyant forces, since it is less dense than oil and water. Once it encounters a caprock, it will spread laterally until it encounters a gap. If there are fault planes in caprock, there is a possibility the  $CO_2$  could migrate along the fault to the surface. Reservoir geometry controls the distribution and migration direction of  $CO_2$ . Formation depth is related to pore pressure, thus impacts the storage capacity and power to inject  $CO_2$ .

Numerous researches have been focusing on the storage site selection, dynamics of  $CO_2$  injection and storage in formation,  $CO_2$  migration in reservoir, and possible  $CO_2$  leak through caprock. House et al.

(2003) simulated CO<sub>2</sub> sequestration in two gas fields in North Sea. The gas reservoirs contain 4-9% CO<sub>2</sub> are high permeability sandstone. The authors created two models to estimate the ultimate storage capacity of the reservoir and assess the risk of reservoir leakage. One model assumed a homogeneous sand with properties of the Utsira sand in the Sleipner area. Another based on a hypothetical faulted anticline with high vertical permeability to fault. The results indicated that geologic complexity is a factor influencing migration of  $CO_2$  in the Miocene reservoir sand and the high vertical permeability faults provide high permeability conduits for the  $CO_2$  leak out of the sequestration reservoir. Akanni et al. (2004) developed material balance model for CO<sub>2</sub> sequestration in depleted gas reservoirs. The storage of CO<sub>2</sub> in dry gas, wet gas, and gas condensate reservoir is compared with depleted oil reservoir. The models calculated higher CO<sub>2</sub> storage volume in gas reservoir than oil reservoir and resulted in higher gas recovery than oil recovery. The developed material balance equations can be utilized to predict CO<sub>2</sub> sequestration volume and survey sequestration operation during injection. p/z plot to forecast the sequestration were also presented. Izgec et al. (2005) used experiment and numerical simulation to investigate the injection of CO<sub>2</sub> in deep saline aquifer. The target formation is Midyat carbonate aquifer located in south-east Turkey. Computerized tomography (CT) was used to monitor porosity and permeability change at different CO<sub>2</sub> injection rates, pressures, and temperatures. The effect of salt concentration on the rock property change was also investigated. The experiment was used to calibrate numerical model created by a finite difference, non-isothermal compositional simulator. Solution and dissolution of carbonates via chemical reaction were considered. Bennion and Bachu (2005) presented the relative permeability of three sandstone and three carbonate formation in the Wabamun Lake area southwest of Edmonton in Alberta, Canada, where four coal-fired power plants which produce large volume of CO<sub>2</sub> are located. The formations are representative of the in-situ temperature, pressure, salinity, porosity and permeability characteristics of deep saline aquifers in on-shore North American sedimentary basins. The impact of CO<sub>2</sub>-water relative permeability on the CO<sub>2</sub> displacing water performance during the CO<sub>2</sub> injection and on ultimate storage capacity was studied. Spiteri et al. (2005) investigated the hysteresis in the relative permeability of the hydrocarbon phase in a two-phase system. The hysteresis is then evaluated to model geological CO<sub>2</sub> sequestration in saline aquifers. In the sequestration, CO<sub>2</sub> is the non-wetting phase and the trapping of the CO<sub>2</sub> is an essential mechanism during lateral and upward migration of CO<sub>2</sub> plume. Sakurai et al. (2006) described the monitoring of saturation change in a CO<sub>2</sub> sequestration in a pilot test in South Lib erty field, Dayton, Texas. The monitoring was through an observation well that was 100 ft away from the injection well. The depth of the target aquifer is 5000 ft with porosity of 32-35% and air permeability of 2500 mD. 1600 metric tons of CO<sub>2</sub> was injected for eleven days. Because of high formation water salinity, along with high porosity, a pulsed neutron tool was selected as the primary log for monitoring saturation changes through change of thermal neutron absorption cross-section. The C/O ratio measurement and dipole acoustic tool were also used to estimate saturation changes. Breakthrough of CO<sub>2</sub> was observed on the third day. At the end of the injection phase, CO<sub>2</sub> saturation of up to 80% was accounted in the observation well. Even four and a half months after the injection experiments, CO<sub>2</sub> saturation of up to 40% was found in the vicinity of the observation and the injection wells, along with temperature anomalies. Schembre-McCabe et al. (2007) researched the mechanistic of CO<sub>2</sub> sequestration in aquifer and depleted hydrocarbon reservoirs using a compositional simulator. The effects of relative permeability and mobility of different phase, damage due to geochemical reaction, stimulation, and nearby injectors were probed. Their study showed that due to limited storage capacity of a reservoir, injectivity declines during the injection stage. This has impact on injector number. Interference among injectors also affects injector performance. Riaz and Tchelepi (2008) analyzed major factors that control the distribution of CO<sub>2</sub> plume by considering the buoyancy and ignoring trapping, dissolution and chemical reaction. Onedimensional model was applied to characterize a pair of shocks moving in opposite direction: top end and bottom end of the plume. Two-dimensional numerical model was used to understand the plume evolution in terms of viscosity ratio and the capillary number. The study found that the buoyancy force causes the

plume to rise up about 500 meter in 700 years. Juanes and MacMinn (2008) proposed a one-dimensional sharp-interface mathematical model for CO<sub>2</sub> migration in saline aquifers. The model accounted for the capillary trapping and gravity override. The CO<sub>2</sub> plume shape was different from previous researches in that the flow of regional groundwater was taken into account. The buoyancy forces lead to gravity override while capillary trapping delays the development of override. Three dimensionless groups: the mobility ratio between  $CO_2$  and water, the gravity number, and trapping coefficient are the key factors that influent shape of  $CO_2$  plume. The capillary trapping coefficient was upscaled from laboratory scale to the basin scale through a formulation based on analytical solution to the model. Nghiem et al. (2009) examined the two most important trapping mechanism for CO<sub>2</sub> storage in saline aquifers: residual gas trapping and solubility trapping. Their study compared how a water injector that is above a CO<sub>2</sub> injector affects the storage of  $CO_2$ . In other words, the formation that is used to store  $CO_2$  lies below the saline aquifer. The results indicated that the low permeability aquifer above the CO<sub>2</sub>-storage formation can increase the storage capacity while high permeability one does not increase CO<sub>2</sub> trapping. Oloruntobi and LaForce (2009) conducted experimental study to investigate the effect of heterogeneity on buoyancy-driven flow in CO<sub>2</sub> storage. Air was used in the study. The study focused on post-injection period when the migration of  $CO_2$  plume in an aquifer is mainly due to the buoyancy drive that results from the density contrast between the injected CO<sub>2</sub> and formation brine. The study showed that air migration through a homogeneous sand pack provided a more uniform front than the heterogeneous sand pack. A high permeability streak in the sand pack reduced the trapping capacity of the entire pack. The presence of a low-permeability layer in the sand pack increased its trapping capacity. The results also depicted that better consolidated sands trapped more air than poorly consolidated sands. Mantilla et al. (2009) used injection data to assess the migration of plume through adapting a probabilistic history matching software originally developed for oil field applications. The software assimilates injection data commonly monitored at active and inactive wells into models for the subsurface aquifer/reservoir. The algorithm vields an ensemble of realizations geologically consistent with the initial model such that the uncertainty in CO<sub>2</sub> plume location can be easily assessed. The updated models obtained after the history matching process detected the presence of the streak, and the subsequent estimation of CO<sub>2</sub> plume location was much more accurate. Murray et al. (2010) summarized a CO<sub>2</sub> sequestration monitoring in a low formation water salinity reservoir in Japan. The targeted reservoir's water salinity is not sufficiently saline so that uncertainty in the Sigma logging approach increases. To overcome this apparent complication, Sigma logging and alternative measurements such as inelastic ratios and neutron porosity were acquired and combined with the open-hole resistivity, neutron-density and magnetic resonance logs to derive a robust interpretation. Liner et al. (2011) used simulation model to evaluate  $CO_2$  sequestration potential in a deep saline aquifer system in Ness County, Kansas. The study showed that a careful simulation study can maximize CO<sub>2</sub> injection rate, minimize existence of free CO<sub>2</sub>, and significantly reduce uncertainty in the safety of CO<sub>2</sub> permanent storage. Amirlatifi et al. (2012) studied the storage of CO<sub>2</sub> in a generic anticline structure. The influence of layer thickness, wavelength and amplitudes at different depths and under different boundary conditions on the maximum CO2 storage amount was investigated. Their 3dimensional model was created by finite element analysis preprocessor. Fluid flow under different geometrical and physical conditions was simulated. Pilisi et al. (2012) investigated the technical feasibility of CO<sub>2</sub> sequestration in deepwater offshore Japan. Sub-seabed formations at water depth of 9000 ft in three sites are selected to conduct reservoir simulation. The study indicated that CO<sub>2</sub> capturing technologies and transportation means are at mature stage and drilling vessels are capable to drill well in such deep water. From the technical point of view, CO<sub>2</sub> sequestration in sub-seabed geological formation is feasible. The three studied sites are appropriate for  $CO_2$  storage. Adebayo et al. (2014) emphasized the importance of CO<sub>2</sub> saturation in aquifer during the sequestration. They determined Archie's saturation exponent for CO<sub>2</sub> sequestration in carbonate reservoirs. They proposed a method that can capture the effect of CO<sub>2</sub>/brine/rock interaction. The method was compared with conventional porous plate method

and it was found that the saturation index differs for both methods largely due to change in pore geometry during CO<sub>2</sub> storage. Park et al. (2015) analyzed the influence of heterogeneous capillary pressure on CO<sub>2</sub> sequestration under different wetting conditions. In their study, numerical simulations are used to estimate the quantity of CO<sub>2</sub> trapped by dissolution, residual, and capillary trapping. Sensitivity analysis is implemented regarding heterogeneity and CO<sub>2</sub>/brine wettability. The research depicted that CO<sub>2</sub> leakage decreases as the permeability variation increases. In strongly water-wet system, the largest portion of stored  $CO_2$  is immobilized. Ajibola et al. (2016) believed the dominant mixing mechanism is convective mixing rather than pure diffusion is important as this controls the timescale over which the carbon dioxidesaturated brine mixes with the unsaturated brine. They used a finite difference reservoir simulator to evaluate the predictions of analytical solutions for stability analysis and growth rate of the fingers of different wavenumbers at different Rayleigh numbers. The effects of density difference, permeability anisotropy and diffusion on fingering behavior were investigated through the dimensionless Rayleigh number. The density difference and the vertical permeability were found to mainly control the degree of instability. Ratnakar et al. (2018) predicted gas solubility in brine solutions for CO<sub>2</sub> capture and sequestration. They extended the Setschenow approach by expressing the solubility in terms of ionic strengths and molar concentrations of each salt. The method also characterized each component (gas, anions and cations) against the experimental measurements. A simple methodology was proposed to predict the impact of types of salts on solubility of CO<sub>2</sub>.

In the CO<sub>2</sub> sequestration process, the injected CO<sub>2</sub> is displacing water from the injection point and is expected to remain in the reservoir. Due to the nature of one phase displacing another phase in porous media, it is noted that different water saturation exists in the CO<sub>2</sub> plume during the displacement. Water distribution in the plume will affect the size of the plume subsurface. Furthermore, the gravitational segregation between CO<sub>2</sub> and water will cause overriding- tonguing during the injection and impact the shape of plume. To better understand the CO<sub>2</sub> movement underground and development of CO<sub>2</sub> plume, it is necessary to take the two-phase flow and gravity force effects into account when evaluating CO<sub>2</sub> displacing water. The purpose of this study is to develop a simple and practical model to predict the evolution of the shape of CO<sub>2</sub> plume in the injection and post-injection period.

### **Model Development**

The displacement of water by injecting  $CO_2$  is not a piston-like process in aquifer. Because water is the wetting phase and  $CO_2$  is the non-wetting phase when two phases flow in reservoir, water occupies the surface of matrix and small pores while CO2 resides in large pores and centers of pores. As  $CO_2$  is injected into aquifer,  $CO_2$  displaces water away from the injector through the fractional flow. At the same time,  $CO_2$  migrates upward due to the buoyancy forces. As a result, in the horizontal direction,  $CO_2$ migration is controlled by relative permeability and capillary pressure. Meanwhile, in the vertical direction, the upward migration of  $CO_2$  is affected by relative permeability, capillary pressure and buoyancy force.

Figure 3 shows a circular aquifer with an injector located in the center. For  $CO_2$  sequestration in aquifer, Figure 3 can represent the sequestration procedure well. The fractional flow can be viewed as  $CO_2$  displacing water away from the injector. Figure 4 illustrates the flow line and pressure distribution in the reservoir. To make the analysis simple, following assumptions are made:

- 1) A circular aquifer with constant height
- 2) Reservoir is homogenous in all rock properties
- 3) The dip angle of formation is zero

- 4) CO<sub>2</sub> and water two-phase flow in reservoir
- 5) Compressibility of water is negligible
- 6) The variation in water densities can be neglected
- 7) Reservoir temperature is constant
- 8) All rock properties do not change during the sequestration.
- 9) Constant water viscosities during the displacement
- 10) Average values are used for CO<sub>2</sub> properties such as viscosity, z-factor, and density when CO<sub>2</sub> flows in reservoir
- 11) Chemical reaction or the dissolution/precipitation can be ignored.



Figure 3. A circular aquifer with an injector located in the center



Figure 4. Reservoir system for CO<sub>2</sub> displacing water away from injector: (left) plan view, (right) lateral view

The flow occurs in two directions: horizontal and vertical flows. We first derive the two-phase flow in horizontal direction. Then vertical flow, or segregation of  $CO_2$  and water in the vertical direction, will be discussed next. In the horizontal direction,  $CO_2$  is displacing water away from the injector. Darcy's equation gives  $CO_2$  (noted as gas in the derivation) and water flow rates

 $q_g = \frac{kk_{rg}}{\mu_g} \frac{\partial(Ap_g)}{\partial r}.$  (1)  $q_{w} = \frac{\bar{kk_{rw}}}{\mu_{w}} \frac{\partial(Ap_{w})}{\partial r}.$ (2) where A = flow area k = reservoir permeability  $k_{rg}$  = relative permeability to gas  $k_{rw}$  = relative permeability to water  $p_g = \text{gas pressure}$  $p_w$  = water pressure  $q_g = \text{gas rate}$  $q_w$  = water rate r = radius from wellbore  $\mu_g = \text{gas viscosity}$  $\mu_w$  = water viscosity In the displacement, CO<sub>2</sub> is the non-wetting phase and water is the wetting phase. Recalling the concept of capillary pressure, we have  $P_c = p_g - p_w.$ where  $P_c$  = capillary pressure Replacing water pressure by gas and capillary pressure Equation (2) becomes  $q_{w} = \frac{kk_{rw}}{\mu_{w}} \frac{\partial [A(p_{g} - p_{c})]}{\partial r}.$ (4) Expressing in pressure gradient, Equations (1) and (4) are changed to  $\frac{\partial(Ap_g)}{\partial r} = \frac{\mu_g}{kk_{rg}} q_g.$ (5)  $\frac{\partial(Ap_g)}{\partial r} - \frac{\partial(AP_c)}{\partial r} = \frac{\mu_w}{kk_{rw}} q_w.$ (6) Subtracting **Equation** (5) from (6) we obtain  $-\frac{\partial (AP_c)}{\partial r} = \frac{\mu_w}{kk_{rw}} q_w - \frac{\mu_g}{kk_{rg}} q_g$ or  $\frac{\partial(AP_c)}{\partial r} = \frac{1}{k} \left( \frac{\mu_g}{k_{rg}} q_g - \frac{\mu_w}{k_{rw}} q_w \right). \tag{7}$ At this stage we can introduce the concepts of total fluid rate and fractional flow, which are defined as  $q_t = q_g + q_w.$  $f_g = \frac{q_g}{q_t}$  and  $f_w = \frac{q_w}{q_t}$ . (9) where

 $q_t =$ total fluid rate

 $f_g$  = gas fraction  $f_w$  = water fraction Substituting Equations (8) and (9) into (7) yields

Substituting Equations (6) and (9) into (7) yields  

$$f_g = \frac{1 + \frac{\partial (AP_C)k_{FW}}{\partial r} \frac{1}{q_{L}\mu_W}}{1 + \frac{k_{FW}\mu_g}{k_{Fg}\mu_W}}.$$
(10)  
Flow area is defined as  
 $A = 2\pi rh.....(11)$   
where  
 $h = \text{reservoir thickness}$   
Substituting Equation (11) into (10) we have  
 $f_g = \frac{1 + \frac{2\pi h k k_{FW}}{q_L \mu_W} \left(\frac{r \partial P_C}{\partial r} + P_C\right)}{1 + \frac{k_{FW}\mu_g}{k_{Fg}\mu_W}}....(12)$   
where  
 $\frac{\partial P_C}{\partial r} = \frac{\partial P_C}{\partial S_g} \frac{\partial S_g}{\partial r}....(13)$   
It should be noted that capillary pressure decreases as radius increases in this case. Therefore  $\frac{\partial P_C}{\partial r}$  is

negative. If capillary pressure is negligible, **Equation** (12) is simplified to

$$f_g = \frac{1}{1 + \frac{k_{rw}\mu_g}{k_{rg}\mu_W}}.$$
 (14)

According to Brooks-Corey (1954 and 1964) capillary pressure model, drainage capillary pressure in the CO<sub>2</sub> sequestration in aquifer can be expressed as

$$P_c = P_d \left(\frac{S_w - S_{wirr}}{1 - S_{wirr}}\right)^{\frac{-1}{\lambda}}.$$
(15)

where

 $S_{wirr}$  = the irreducible water saturation

 $P_d$  = the threshold pressure.

Brooks and Corey (1954 and 1964) related the parameter  $\lambda$  to the distribution of pore sizes. For narrow distributions,  $\lambda$  is greater than 2; for wide distributions,  $\lambda$  is less than 2. **Equation** (12) indicates that gas fraction is a function of location and injection volume (or injection time if injection rate is kept constant). Now we derive the continuity equation of radial fractional flow. Considering the CO<sub>2</sub> displacing water, material balance equation provides that the mass change in a control volume for a time period can be shown as **Figure** 5. Since the flow direction is from injector to aquifer outer boundary, we define the center of wellbore as the start point, where *r*=0, and aquifer outer boundary as the end point where *r*=*r*<sub>e</sub>. Therefore, in the dimensionless analysis, dimensionless radius can be defined as,  $r_D=r/r_e$ , the start point has  $r_D=0$ , and the end point has  $r_D=1$ . Material balance gives

$$\left[\left(q_{g}\rho_{g}\right)_{r}-\left(q_{g}\rho_{g}\right)_{r+\Delta r}\right]\Delta t=\pi h\left[\left(r+\Delta r\right)^{2}-r^{2}\right]\varphi\left[\left(S_{g}\rho_{g}\right)^{t+\Delta t}-\left(S_{g}\rho_{g}\right)^{t}\right].$$
(16)

where

 $S_g = gas saturation$ 

 $\rho_g = \text{gas density}$ 

 $\Delta t = \text{time period}$ 

t = time

 $\Delta r$  = radius incremental

*r*= radius

 $r_e$ = distance between reservoir outer boundary to wellbore

 $\varphi = \text{porosity}$ 



Figure 5. A control volume in a circular aquifer with an injector located in the center.

Simplifying Equation (16) we have
$\left[\left(q_g\rho_g\right)_r - \left(q_g\rho_g\right)_{r+\Delta r}\right]\Delta t = \pi h [2r\Delta r + (\Delta r)^2]\varphi \left[\left(S_g\rho_g\right)^{t+\Delta t} - \left(S_g\rho_g\right)^t\right].$ (17)
As $\Delta r \to 0$ and $\Delta t \to 0$ , we have
$2r\Delta r + (\Delta r)^2 \approx 2r\Delta r.$ (18)
Equation (17) becomes partial differential equation
$-\frac{\partial(q_g\rho_g)}{\partial r} = 2r\pi h\varphi \frac{\partial(S_g\rho_g)}{\partial t}.$ (19)
If an average gas density is used and is assumed constant, we have
$-\frac{\partial q_g}{\partial r} = 2r\pi h\varphi \frac{\partial S_g}{\partial t}.$ (20)
Substituting Equation (9) into (20) gives
$-\frac{\partial(f_g q_t)}{\partial r} = 2r\pi h \varphi \frac{\partial S_g}{\partial t}.$ (21)
If CO <sub>2</sub> injection rate is constant and an average CO <sub>2</sub> density at aquifer temperature and average pressure
(average of injection pressure and aquifer outer boundary pressure) is used to represent CO <sub>2</sub> density in
aquifer, we have a constant total fluid rate. Equation (21) can be simplified to
$-\frac{\partial f_g}{\partial t_g} = \frac{2r\pi h\varphi}{2}\frac{\partial S_g}{\partial t_g}.$ (22)
$\partial r = q_t  \partial t$
Since water fraction is function of water saturation, $J_g(S_g)$ , applying chain rule to partial differential
equation results in
$-\frac{dg}{ds}\frac{dsg}{dr} = \frac{2\pi nn\phi}{a}\frac{dsg}{dt}.$ (23)
Observing that water saturation is function of time t and position r we can express
$\partial S_g d_{L_1} = \partial S_g d_{L_2}$ (24)
$aS_g = \frac{1}{\partial t}at + \frac{1}{\partial r}ar.$ (24)
The fact that at the displacement front the $CO_2$ saturation is constant provides us a boundary condition.
$dS_g = \frac{\partial S_g}{\partial t}dt + \frac{\partial S_g}{\partial r}dr = 0$
or
$\frac{\partial s_g}{\partial t} = -\frac{\partial s_g}{\partial t} \frac{dt}{dt}.$ (25)
$\frac{\partial r}{\partial t} = \frac{\partial t}{\partial t} \frac{dr}{dr}$ Substituting Equation (25) into (23) yields

$\frac{df_g}{dt} dt = \frac{2r\pi\hbar\varphi}{dt} dt \qquad (26)$	5)
$dS_g \qquad q_t$	
Integrating Equation (26) yields an equation for displacement front position, $r_f$ .	
$r_f^2 = \frac{tq_t}{\pi h\varphi} \left(\frac{df_g}{dS_g}\right)_f \dots \dots$	')

where

 $r_f$  = displacement front position in radial aquifer There are two solutions to **Equation** (27), which are

$$r_f = \pm \sqrt{\frac{tq_t}{\pi h \varphi} \left(\frac{df_g}{dS_g}\right)_f}.$$
(28)

Obviously only one solution is correct to match with the physical phenomenon. Considering the fact that location can only be positive, therefore we can eliminate the solution

$$r_f = -\sqrt{\frac{tq_t}{\pi h\varphi} \left(\frac{df_g}{dS_g}\right)_f}.$$
(29)

Therefore, the correct solution is

$$r_f = \sqrt{\frac{tq_t}{\pi h\varphi} \left(\frac{df_g}{dS_g}\right)_f}...(30)$$

For any  $CO_2$  saturation that is behind the displacement front,  $S_g$ , the position can be calculated by

$$r_{Sg} = \sqrt{\frac{tq_t}{\pi h\varphi} \left(\frac{df_g}{dS_g}\right)_{Sg}}....(31)$$

where

 $r_{Sg}$  = position of any CO<sub>2</sub> saturation that is behind the displacement front

Water saturation ahead of the displacement front is the initial water saturation in reservoir. For aquifer, initial water saturation can be assumed 100%. CO<sub>2</sub> saturation ahead of displacement front is 0. Therefore, locations of different  $CO_2$  saturations can be estimated accordingly based on **Equations** (31). Figure 7 shows the plot of  $CO_2$  saturation versus radius at different injection times assuming vertical segregation of CO<sub>2</sub> and water can be ignored.

Because of the density difference between  $CO_2$  and water. Segregation between  $CO_2$  and water occurs when  $CO_2$  mixes with water in reservoir. In the segregation, water flows downward while  $CO_2$ flows upward. Considering a control volume around the well (or a ring-shape of reservoir volume) at a radius of r, material balance gives

 $[(q_{wV}\rho_w)_{z+\Delta z} - (q_{wV}\rho_w)_z]\tilde{\Delta t} = A\Delta z\varphi[(S_w\rho_w)^{t+\Delta t} - (S_w\rho_w)^t].$ (32) where  $q_{WV}$  = water flowrate in vertical direction  $\rho_w$  = water density A = flow area of control volume when vertical flow occurs  $\Delta z$  = incremental in vertical direction  $S_w$  = water saturation As  $\Delta z \rightarrow 0$  and  $\Delta t \rightarrow 0$ , we have Equation (32) becomes partial differential equation  $\frac{\partial(\bar{q}_{WV}\rho_W)}{\partial z} = A\varphi \frac{\partial(S_W\rho_W)}{\partial t}.$ (33)

Assuming constant water density, Equation (33) can be simplified to

$\frac{\partial(q_{wV})}{\partial z} = A\varphi \frac{\partial(S_w)}{\partial t}.$ (34)
The flow of fluid in vertical direction is controlled by buoyancy forces, capillary pressure, and vertical permeability. While the buoyancy force facilitates the water downward migration, the effect of capillary pressure depends on the wettability. In CO <sub>2</sub> -water system, water is wetting phase and CO <sub>2</sub> is non-wetting
phase. Therefore, capillary pressure facilitates water flow as well. Darcy law gives
$q_{wV} = A \frac{k_V k_{rw}}{\mu_w} \frac{\Delta z (\rho_w - \rho_g) + P_c}{\Delta z}.$ (35)
where
$k_V =$ vertical permeability
Substituting Equation (35) into (34) yields
$\frac{k_V}{\mu_W}(\rho_W - \rho_g)\partial(k_{rW}) + \partial\left(\frac{P_c}{\Delta z}\right) = \alpha \partial(S_W) $ (26)
$\frac{\partial z}{\partial z} + \frac{\partial z}{\partial z} - \frac{\varphi}{\partial t} \dots $
If capillary pressure is negligible, <b>Equation</b> (36) can be simplified to $\frac{k_{\rm E}}{k_{\rm E}}$
$\frac{k \nu}{\mu_w} (\rho_w - \rho_g) \partial(k_{rw}) = \alpha \frac{\partial(S_w)}{\partial k_{rw}} $ (37)
$\partial z = -\varphi_{\partial t}$
$\begin{bmatrix} f \\ f $
$\left[\left(q_{gV}\rho_{g}\right)_{z}-\left(q_{gV}\rho_{g}\right)_{z+\Delta z}\right]\Delta t=A\Delta z\varphi\left[\left(S_{g}\rho_{g}\right)-\left(S_{g}\rho_{g}\right)\right].$ (38)
where
$q_{gV}$ = gas flowrate in vertical direction
As $\Delta z \to 0$ and $\Delta t \to 0$ , we have
Equation (38) becomes partial differential equation
$-\frac{\partial(q_{gV}\rho_g)}{\partial t} = A\varphi \frac{\partial(s_g\rho_g)}{\partial t}.$ (39)
Assuming average gas density is used <b>Equation (39)</b> can be simplified to
$-\frac{\partial(q_{gV})}{\partial z} = A\varphi \frac{\partial(S_g)}{\partial t}.$ (40)
Buoyancy force facilitates the gas upward migration while capillary pressure has the opposite effect.
Darcy law gives
$q_{gV} = A \frac{k_V k_{rg}}{\mu_g} \frac{\Delta z (\rho_w - \rho_g) - P_c}{\Delta z}.$ (41)
Substituting Equation (41) into (40) yields
$-\frac{\frac{k_V}{\mu_g}(\rho_w - \rho_g)\partial(k_{rg})}{\mu_g} + \frac{\partial(\frac{P_c}{\Delta z})}{\mu_g} = \omega \frac{\partial(S_g)}{\mu_g}.$ (42)
$-\frac{\frac{k_V}{\mu_g}(\rho_W - \rho_g)\partial(k_{rg})}{\partial z} + \frac{\partial(\frac{P_c}{\Delta z})}{\partial z} = \varphi \frac{\partial(S_g)}{\partial t}.$ (42) If capillary pressure is pegligible. Equation (42) can be simplified to
$-\frac{\frac{k_V}{\mu_g}(\rho_w - \rho_g)\partial(k_{rg})}{\partial z} + \frac{\partial(\frac{P_c}{\Delta z})}{\partial z} = \varphi \frac{\partial(s_g)}{\partial t}.$ (42) If capillary pressure is negligible, <b>Equation (42)</b> can be simplified to
$-\frac{\frac{k_V}{\mu_g}(\rho_w - \rho_g)\partial(k_{rg})}{\partial z} + \frac{\partial(\frac{P_c}{\Delta z})}{\partial z} = \varphi \frac{\partial(s_g)}{\partial t}.$ (42) If capillary pressure is negligible, <b>Equation (42)</b> can be simplified to $-\frac{\frac{k_V}{\mu_g}(\rho_w - \rho_g)\partial(k_{rg})}{\frac{\mu_g}{\partial t}} = \varphi \frac{\partial(s_g)}{\partial t}.$ (43)
$-\frac{\frac{k_V}{\mu_g}(\rho_w - \rho_g)\partial(k_{rg})}{\partial z} + \frac{\partial(\frac{P_c}{\Delta z})}{\partial z} = \varphi \frac{\partial(S_g)}{\partial t}.$ (42) If capillary pressure is negligible, <b>Equation (42)</b> can be simplified to $-\frac{\frac{k_V}{\mu_g}(\rho_w - \rho_g)\partial(k_{rg})}{\partial z} = \varphi \frac{\partial(S_g)}{\partial t}.$ (43) Coupling <b>Equation (31)</b> for horizontal flow with <b>Equations (37)</b> and <b>(43)</b> for vertical COs water

segregation,  $CO_2$  saturation distribution in the aquifer at different injection times or post-injection periods can be calculated.

## **Case Study**

A case study was conducted to illustrate the analysis of  $CO_2$  displacing water in aquifer. The input data are shown in **Tables** 3. **Table** 4 is the relative permeabilities versus gas and water saturations. Figure 6 shows the capillary pressures of  $CO_2$ -water system versus water saturation used in the case study. It is a drainage curve that simulates non-wetting phase ( $CO_2$ ) displacing water (wetting phase).

CO <sub>2</sub> injection rate at standard conditions	Ξ	20,000	Mscf/day
Aquifer radius	=	10,000	ft
Aquifer thickness	=	100	ft
Initial water saturation	Ξ	100%	
Irreducible water saturation	Ξ	20.0%	
Residual CO <sub>2</sub> saturation	=	25.0%	
CO <sub>2</sub> viscosity	=	0.02	ср
Water viscosity	Ξ	1.0	ср
Porosity	Ξ	20.0%	
Reservoir horizontal permeability	Ξ	200	mD
		20, 100, or	
Reservoir vertical permeability	=	200	mD
Reservoir temperature	Ξ	150	°F
Average pressure in injection calculation	Π	3500	psia
Aquifer porosity	Ξ	0.25	
CO <sub>2</sub> compressibility factor at reservoir			
conditions	=	1	

# Table 3. The input data for $CO_2$ displacing water in an aquifer system

**Table 4.** The relative permeabilities versus water saturation

No.	Sw	Sg	krw	krg
	0	1	0	0.5
1	0.2	0.8	0	0.5
2	0.209	0.791	0.0001	0.4838
3	0.218	0.782	0.0005	0.4678
4	0.227	0.773	0.0012	0.4522
5	0.236	0.764	0.0021	0.4368
6	0.245	0.755	0.0033	0.4216
7	0.254	0.746	0.0048	0.4067
8	0.263	0.737	0.0065	0.3921
9	0.272	0.728	0.0085	0.3778
10	0.281	0.719	0.0108	0.3637
11	0.290	0.710	0.0134	0.3499
12	0.299	0.701	0.0162	0.3364
13	0.308	0.692	0.0192	0.3231
14	0.317	0.683	0.0226	0.3101
15	0.326	0.674	0.0262	0.2974
16	0.335	0.665	0.0300	0.2849
17	0.344	0.656	0.0342	0.2727
18	0.353	0.647	0.0386	0.2608

19	0.362	0.638	0.0433	0.2491
20	0.371	0.629	0.0482	0.2377
21	0.380	0.620	0.0534	0.2266
22	0.389	0.611	0.0589	0.2157
23	0.398	0.602	0.0646	0.2051
24	0.407	0.593	0.0706	0.1948
25	0.416	0.584	0.0769	0.1847
26	0.425	0.575	0.0835	0.1749
27	0.434	0.566	0.0903	0.1654
28	0.443	0.557	0.0974	0.1561
29	0.452	0.548	0.1047	0.1471
30	0.461	0.539	0.1123	0.1384
31	0.470	0.530	0.1202	0.1299
32	0.479	0.521	0.1283	0.1217
33	0.488	0.512	0.1368	0.1138
34	0.497	0.503	0.1454	0.1061
35	0.506	0.494	0.1544	0.0987
36	0.515	0.485	0.1636	0.0916
37	0.524	0.476	0.1731	0.0847
38	0.533	0.467	0.1828	0.0781
39	0.542	0.458	0.1928	0.0718
40	0.551	0.449	0.2031	0.0657
41	0.560	0.440	0.2137	0.0600
42	0.569	0.431	0.2245	0.0544
43	0.578	0.422	0.2356	0.0492
44	0.587	0.413	0.2469	0.0442
45	0.596	0.404	0.2586	0.0395
46	0.604	0.396	0.2704	0.0350
47	0.613	0.387	0.2826	0.0308
48	0.622	0.378	0.2950	0.0269
49	0.631	0.369	0.3077	0.0232
50	0.640	0.360	0.3207	0.0198
51	0.649	0.351	0.3339	0.0167
52	0.658	0.342	0.3474	0.0139
53	0.667	0.333	0.3611	0.0113
54	0.676	0.324	0.3751	0.0090
55	0.685	0.315	0.3894	0.0069
56	0.694	0.306	0.4040	0.0051
57	0.703	0.297	0.4188	0.0036
58	0.712	0.288	0.4339	0.0023
59	0.721	0.279	0.4493	0.0014
60	0.730	0.270	0.4649	0.0006

61	0.739	0.261	0.4808	0.0002
62	0.748	0.252	0.4969	0.0000
63	0.757	0.243	0.5000	0.0000
64	0.766	0.234	0.5000	0.0000
65	0.775	0.225	0.5000	0.0000
66	0.784	0.216	0.5000	0.0000
67	0.793	0.207	0.5000	0.0000
68	0.802	0.198	0.5000	0.0000
69	0.811	0.189	0.5000	0.0000
70	0.820	0.180	0.5000	0.0000
71	0.829	0.171	0.5000	0.0000
72	0.838	0.162	0.5000	0.0000
73	0.847	0.153	0.5000	0.0000
74	0.856	0.144	0.5000	0.0000
75	0.865	0.135	0.5000	0.0000
76	0.874	0.126	0.5000	0.0000
77	0.883	0.117	0.5000	0.0000
78	0.892	0.108	0.5000	0.0000
79	0.901	0.099	0.5000	0.0000
80	0.910	0.090	0.5000	0.0000
81	0.919	0.081	0.5000	0.0000
82	0.928	0.072	0.5000	0.0000
83	0.937	0.063	0.5000	0.0000
84	0.946	0.054	0.5000	0.0000
85	0.955	0.045	0.5000	0.0000
86	0.964	0.036	0.5000	0.0000
87	0.973	0.027	0.5000	0.0000
88	0.982	0.018	0.5000	0.0000
89	0.991	0.009	0.5000	0.0000
90	1.000	0.000	0.5000	0.0000



Figure 6. Capillary pressure versus water saturation for the CO<sub>2</sub> displacing water in aquifer during injection

If vertical segregation is ignored,  $CO_2$  saturation at any radius is a unique number and is constant at different aquifer elevations. **Figure 7** illustrates the  $CO_2$  saturation distribution along the radius of aquifer at different injection times. In this case,  $CO_2$  saturation at the displacement front is 0.342. Ahead of the displacement front, water saturation is 100%, or  $CO_2$  saturation is 0.0%. The location of displacement front at different injection time can be found in **Table 5**. **Figure 7** and **Table 5** indicate that movement of  $CO_2$  saturation away from the well is not linear to the injection time. The movement slows down because more pore volumes needed to be filled as radius increases.



Figure 7. Change of CO<sub>2</sub> saturation in aquifer during different injection times

Injection													
time (days)	1	2	5	10	20	50	100	200	500	1000	2000	5000	10000
Displacement													
front													
Location (ft)	89	126	199	281	398	629	889	1257	1988	2811	3976	6286	8890

Table 5. Location of displacement front at different injection times

 $CO_2$ -water segregation causes  $CO_2$  to migrate to the top of aquifer. **Figure 8** depicts the variation of  $CO_2$  saturation from the bottom of aquifer to the top of aquifer at different radii. It can be seen that near the wellbore the segregation is more complete than away from the wellbore, which is indicated by wide ranges of  $CO_2$  saturation between the bottom of aquifer (with aquifer elevation of 0 ft) and the top of aquifer (with aquifer elevation of 100 ft). The segregation leads to the overriding of  $CO_2$  on water during the injection and for post-injection period. Increasing injection rate can reduce the overriding effect temporary, but segregation, the difference between the buoyancy force is balanced by the capillary force. At the beginning of the segregation, the difference between the buoyancy force and capillary force is the highest. The force difference that causes the  $CO_2$  vertical-migration is the largest. As gas migration continues, capillary force increases. Therefore, the segregation speed is highest at the beginning and then slows down as time goes on. Once the buoyancy force is equal to capillary force,  $CO_2$ -water system reaches a dynamic equilibrium status and  $CO_2$  will remain in location until the equilibrium is broken.



Figure 8. Change of CO<sub>2</sub> saturation along aquifer elevation at different radii

Anisotropy of rock properties is often observed in aquifer. For most aquifers vertical permeability is less than horizontal permeability, which alleviates the  $CO_2$  vertical migration. In this study, the effect of vertical/horizontal permeability ratio on the  $CO_2$  overriding on water is investigated. Three cases with vertical/horizontal permeability ratio of 1.0, 0.5 and 0.1 are compared to analyze how low vertical permeability delays the  $CO_2$  upward migration. One should note that low permeability also means high capillary pressure, which also limits the upward flow of  $CO_2$ . **Figure 9** shows the  $CO_2$  saturation distribution along aquifer elevation for different vertical/horizontal permeability ratios at two radii. It clearly shows that low vertical/horizontal permeability ratio.



Figure 9. Effect of permeability anisotropy on CO2-water segregation at same radius

# Conclusions

Following conclusions can be drawn upon this study.

- A model coupling CO<sub>2</sub>-water fractional flow in horizontal direction with CO<sub>2</sub>water segregation in vertical direction has been proposed. It can be used to simulate the performance of CO<sub>2</sub> sequestration in aquifer.
- The model indicates that CO<sub>2</sub> displacing water is not a piston-like displacement. CO<sub>2</sub> saturation at the displacement front is a function of fluid and rock properties. Behind the CO<sub>2</sub> front, CO<sub>2</sub> saturation increases gradually, and water saturation decreases accordingly until water saturation reaches irreducible water saturation.
- The propagation of CO<sub>2</sub> front is not linear to the injection time. Its movement slows down as it moves away from injection well.
- For same radius, CO<sub>2</sub>-water segregation causes higher CO<sub>2</sub> saturation at top of aquifer and low at the bottom of aquifer.
- Low vertical/horizontal permeability ratio can delay CO<sub>2</sub>-water segregation, leads to a less severity of CO<sub>2</sub> overriding.

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### Nomenclature

A = flow area

- A = flow area of control volume when vertical flow occurs
- $f_g = \text{gas fraction}$
- $f_w$  = water fraction
- h = reservoir thickness
- k = reservoir permeability
- $k_{rg}$  = relative permeability to gas
- $k_{rw}$  = relative permeability to water
- $k_V$  = vertical permeability
- $P_c$  = capillary pressure
- $P_d$  = the threshold pressure.
- $p_g = \text{gas pressure}$
- $p_w$  = water pressure
- $q_g = \text{gas rate}$
- $q_{gV}$  = gas flowrate in vertical direction
- $q_t$  = total fluid rate
- $q_w$  = water rate
- $q_{wV}$  = water flowrate in vertical direction
- r = radius from wellbore
- $r_e$  = distance between reservoir outer boundary to wellbore
- $r_f$  = displacement front position in radial aquifer
- $r_{Sg}$  = position of any CO<sub>2</sub> saturation that is behind the displacement front
- $S_g = gas saturation$
- $S_w$  = water saturation
- $S_{wirr}$  = irreducible water saturation
- t = time
- $\mu_g = \text{gas viscosity}$
- $\mu_w$  = water viscosity
- $\rho_g = \text{gas density}$
- $\rho_w$  = water density
- $\varphi = \text{porosity}$
- $\Delta t = \text{time period}$
- $\Delta r$  = radius incremental
- $\Delta z$ = incremental in vertical direction

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